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Sustainability Assessment of the Utilization of CO₂ in a Dielectric Barrier Discharge Reactor Powered by Photovoltaic Energy

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Abstract: The direct activation of diluted CO_2 in argon was studied in a co-axial dielectric barrier discharge (DBD) reactor powered by photovoltaic energy. The influence of the initial CO_2 and argon concentration on the CO_2 decomposition to form CO was investigated using a copper-based catalyst in the discharge zone. It was observed that the CO_2 conversion was higher at lower CO_2 concentrations. The presence of the diluent gas (argon) was also studied and it was observed how it has a high influence on the decomposition of CO_2 , improving the conversion at high argon concentrations. At the highest observed energy efficiency (1.7%), the CO_2 conversion obtained was 40.2%. It was observed that a way to enhance the sustainability of the process was to use photovoltaic energy. Taking into account a life cycle assessment approach (LCA), it was estimated that within the best-case scenario, it would be feasible to counterbalance 97% of the CO_2 emissions related to the process.

Keywords: non-thermal plasma; CO₂ conversion; DBD; carbon capture and utilization (CCU)



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1. Introduction

Nowadays, it is a priority to find and develop technologies to deal with CO_2 emissions in order to mitigate global warming [1]. New alternatives for carbon capture technologies have been studied and assessed [2–6], but the problem of which is the best alternative for CO_2 as a feedstock is still under study [7–10]. That is why CO_2 conversion to valuable products appears to be a great possibility; this leads to the reduction in fossil fuel dependence and global warming too.

The direct use of CO_2 has a great variety of applications, such as its relevance within soft drinks, to the scope of supercritical CO_2 as a solvent, and in other industrial areas, for instance, welding and refrigeration [7]. Direct dissociation of CO_2 and conversion into other value-added fuels and chemicals provides a potential route for efficient reduction in CO_2 emissions. Variable progress has been made to convert CO_2 into other value-added chemicals, such as CO_2 hydrogenation for the synthesis of methane, methanol, formaldehyde, etc. [8–10].

CO₂ conversion, decomposition, or activation into valuable products have been widely studied through thermo-chemical [11], photocatalysis [12–17], biological [18] or electrochemical [19,20] pathways, although these routes are not efficient from an energy perspective. Alternatively, non-thermal plasma (NTP) can operate under ambient conditions, so it can offer high energetic electrons which are able to initiate a highly endothermic chemical reaction under ambient temperature [21]. Different NTP approaches have been broadly studied such as corona discharge [22,23], gliding arc [24,25], microwave discharge [26,27], and dielectric barrier discharge (DBD) [28–30]. The latter has drawn more attention for the wide range of applications and its operability. Among several applications for the DBD approach, it has been successfully applied for volatile organic compounds treatment [31], wastewater treatment [32,33], reforming reaction [34], and methanol production [10,11,19,35]. NTP can operate at room temperature and atmospheric pressure while generating highly active electrons, with mean electron energy between 1 and 10 eV. This electron energy is the optimum range in order to activate molecular and atomic species and break chemical bonds [36]. NTP technologies are advantageous over thermal processes as reaction rates are higher and a steady state is achieved much faster [22]. On the other hand, the DBD reactor has the capacity to produce highly energetic electrons and uniform distribution discharges and is efficient in initiating chemical reactions under room conditions [37–39]. The combination of NTP with the main advantages of DBD reactors enables this system to be an ideal candidate for CO_2 utilization.

Therefore, NTP has high potential as an efficient CO_2 utilization process, as it can overcome the stability of CO_2 without the need for the high temperatures required in thermal catalytic processes. This facilitates quick start-up and shut-down, a promising feature that enables plasma technology powered by renewable energy to act as an efficient chemical energy storage [40].

In this work, the study of the conversion of CO_2 into CO through a DBD plasma reactor was developed by a sensitivity analysis for the CO_2 flowrate and the concentration of Ar that was used as a diluent; then, a sustainability assessment was considered for the suitability of using photovoltaic energy as an energy source to power the DBD plasma reactor.

2. Materials and Methods

2.1. Materials

The following reactants were used: Argon (Ar) (Carburos Metálicos; 99.9997%), Carbon dioxide (CO₂) (Carburos Metálicos; 99.999%) and the copper/zinc-based catalyst (Alfa Aesar; copper-based catalyst (pellets, 5.4 mm \times 3.6 mm)) with a composition of Al₂O₃ 10%, CuO 64%, MgO 2% and ZnO 24%.

2.2. Experimental Setup

Figure 1 shows the experimental setup used in this work. Briefly, a cylindrical quartz tube, length: 250 mm; outer diameter: 25 mm; inner diameter: 24 mm, (Vidrasa, Ripollet, Barcelona, Spain) was used to generate the plasma discharge. A copper rod (Broncesval, Ripollet, Barelona, Spain) with an outer diameter of 5 mm was used as the internal electrode (anode) and the outer electrode (cathode) was a copper mesh of 160 mm, which was wrapped around the quartz tube, as shown in Figure 1. The resulting discharge gap was 10 mm. An amount of 40 g catalyst sandwiched between quartz wool was placed in the discharge area zone. NTP was generated by an AC high voltage power supply with a frequency, duty cycle, and voltage controller: Plasma Drive PVM500/DIDRIVE10 (Amazing1, Mont Vernon, NH, USA). The AC high-voltage power supply was connected to a photovoltaic (PV) system in order to provide the electricity needed to carry out the experimental tests. The PV system, which consisted of one solar panel Tecnosun 150 W, was placed outside to receive sunlight, using two batteries Rolls 6 V and 480 Ah and CC/AC inverter model Victron Energy Phoenix 12/1200. All the CC system worked at 12 V and was controlled by a regulator STECA PR 2020 regulator. In this work, the necessary energy to run the tests was provided directly from the PV panels. The energy consumption was continuously determined with a wattmeter. A multi-channel oscilloscope Promax 0D-610 100 MHz (Promax, L'Hospitalet de Llobregat, Barcelona, Spain) was used to monitor the voltage, frequency and current intensity during the experiences. The amount of CO₂ and Ar in the reactor was controlled with 2 flowmeters EL-FLOW® (Bronkhorst, Nijverheidsstraat, Ruurlo, The Netherlands).



Figure 1. DBD system includes the photovoltaic power generation (solar panel (A), batteries (B), inverter (C), solar regulator (D)), electric signal amplifier (E), oscilloscope (F), volumetric flowmeters (G), copper rod inner electrode (H), wattmeter (I), copper mesh outer electrode (J) and PTFE reactor ends (K). The DBD reactor (L) is filled with catalyst pellets.

2.3. CO₂ Conversion into CO Experiments

All the experiments were performed at maximum voltage (32 kV), minimum frequency (20 kHz) and minimum duty changing flowrates of CO_2 and Ar. All the volumetric flowrates used in this study are given under normal conditions. The power used in all experiments was around 50 W. The total time for each experiment was 30 s once the steady state was reached. Samples were collected in Tedlar bags for further analysis. The samples were analyzed with a gas chromatograph (Agilent 7020A, Santa Clara, CA, USA) equipped with an Agilent HP-PLOT Molesieve 19095P-MS0 column and a thermal conductivity detector (TCD). The injection, oven, and detector temperatures were 100 °C, 60 °C, and 250 °C, respectively. All the experiments were carried out in triplicate and the average values are shown in the figures with the standard deviation as error bars.

2.4. Environmental Sustainability Assessment

The environmental sustainability of the process using PV energy was studied by calculating the carbon footprint. The carbon footprint was assessed with the software SimaPro 9.0 (Amersfoort, The Netherlands) using a Life Cycle Assessment (LCA) approach. For the LCA, the Ecoinvent v3.6 database and the ILCD 2011 Midpoint methodology were used considering two different scenarios. The first scenario considers the supply of electricity from the electrical network and the second scenario considers the electricity provided by a PV system.

To determine the percentage of the CO_2 emissions compensated (EC_{co2}), Equation (1) was used.

$$EC_{co2} (\%) = \frac{Con_{CO2} + P_{CO}}{E_{electricity}} \times 100$$
(1)

where Con_{CO2} (g CO_2 -eq/h) was the CO_2 converted into CO in the DBD reactor and P_{CO} (kg CO_2 -eq/h) was the CO_2 emission that is related to the industrial production of the generated CO with a conventional process such hydrocarbon reforming. Finally, $E_{electricity}$ is the emission of CO_2 (g CO_2 -eq/h) produced by the electrical consumption, which was

measured with a wattmeter. The emission factors to produce electricity (from the electrical network or PV systems) and for the industrial production of CO were extracted from Ecoinvent.

3. Results and Discussion

3.1. Influence of CO₂ Flowrate

Low CO_2 rates of decomposition have been found in the literature using pure CO_2 . whereas the use of Ar has been reported to be a good option due to the use of a diluent gas (gas to be ionized) increases CO_2 conversion. When Ar is used, the main reaction pathway in the NTP zone involves Ar excitation and charge/energy transfer from excited Ar atoms to CO_2 molecules [41]. Mei et al. [28] reported that Ar dilution is beneficial over other gases such as N_2 and He. They observed that CO_2 conversion increases with increasing Ar concentrations. The reason is that by decreasing the dielectric strength of the gas mixture, more energy is available for CO_2 molecules in the discharge. In addition, the available electrons may excite CO₂ molecules contributing to reaching higher conversions due to the lower number of CO₂ molecules with respect to the Ar atoms. Mei et al. [28] also reported selectivities for CO₂ conversion into CO in high Ar concentrations between 90–98% that increased at increasing Ar concentration. For that reason, Ar was used as diluent gas in this work. At the studied conditions, when voltage was applied between the two electrodes in the DBD reactor, NTP discharges were produced within the plasma reactor, as Figure 2 shows. It was observed randomized discharge points within the DBD plasma reactor that continuously changed. The NTP was fast generated or stopped by switching on or off the AC high voltage power supply, quickly reaching the steady state as suggested by Pou et al. [40].



Figure 2. Picture of the NTP discharges across the DBD plasma reactor.

Two important outcomes to check the performance of the reaction are the CO_2 conversion and the energy efficiency (n). The CO_2 conversion (%) was determined with the following equation:

$$CO_2 \text{ Conversion} (\%) = \frac{C_{CO2,in} - C_{CO2,out}}{C_{CO2,in}} \times 100$$
(2)

where $C_{CO2,out}$ is the molar concentration (%) of CO_2 in the outlet stream and $C_{CO2,in}$ is the molar concentration (%) of CO_2 in the inlet stream of the DBD reactor. From the chromatography analysis, it was found that a selectivity of almost 100% for CO was reached for all the tests (no other peaks different from CO or CO_2 were observed). From these results, it was considered that CO was the major product and the stoichiometric conversion of CO_2 into CO was achieved. So, the conversion and yield obtained were almost the same. This point has also been reported in previous works [28,40] where selectivities to CO higher than 95% were observed. Carbon deposition over the catalyst was not detected after CO_2 decomposition with NTP. It must be highlighted that the CO selectivity did not present any dependence on other parameters studied in this work, such as CO_2 or Ar flowrates. Other works shown in Table 1 reported selectivities ranging from 48 to 96%. Therefore, the influence of plasma operational parameters on the NTP process will only be discussed in

terms of CO₂ conversion and energy efficiency. The energy efficiency of the NTP process was calculated using the following equation:

$$\eta(\%) = \frac{F_{CO2} \cdot CO_2 Conversion \cdot \Delta H_{CO}}{60 \times 22.4 \cdot P_E} \times 100$$
(3)

where F_{CO2} is the volumetric flowrate of CO_2 in the inlet stream (mL/min), ΔH_{CO} is the reaction enthalpy of pure CO_2 decomposition into CO (283.1 kJ/mol [42]) and P_E is the electric power used to carry out the reaction (W).

Flowrate (mL/min)	Diluent Gas	Power (W)	Packed Material	CO ₂ Conv. (%)	CO Select. (%)	CO Yield (%)	η (%)	Ref
30	Ar	2.4	Glass beads	19.5	86	16.8	17.0	[39]
150	-	55	Molecular sieves 5A	25	63	15.8	-	[44]
60	-	28	BaTiO ₃	38.3	-	-	17.0	[45]
50	-	50	BaTiO ₃	28	96	26.9	7.2	[46]
30	-	2.4	5% ZnO + g-C3N4	12	70	8.4	31.1	[16]
30	-	2.2	15% CuO/Al ₂ O ₃	15.7	48	7.5	45.2	[43]
30	-	2.2	15% CuO/CeAl	13.5	59	7.8	38.9	[43]
9	Ar	50	$CuO/ZnO/Al_2O_3$	40.2	>99	40.2	1.7	This work

 Table 1. Comparison of DBD plasma-assisted conversion of CO2.

The effect of the CO₂ flowrate on the CO₂ conversion and energy efficiency was evaluated. In Figure 3, it can be observed how the CO₂ conversion decreased by increasing the CO₂ flowrate in the inlet. According to previous research [40], it has been proved that the generated NTP is more uniform at high Ar concentrations. In addition, at lower CO₂ flowrates, the residence time in the reactor was higher and this would increase the probability that a NTP discharge could affect a CO₂ molecule. This behavior has also been described using a fluidized bed NTP reactor with Cu/ γ -Al₂O₃ powder-based catalyst [40]. The maximum CO₂ conversion in this work was 74.2% and it was obtained at the lowest CO₂ flowrate. This conversion was much higher than the one obtained by Ray et al. (maximum conversion of 15.7%) utilizing a packed DBD plasma reactor with Ni and Cu oxide supported γ -Al₂O₃ as catalyst [43]. Other works (Table 1) reported conversions ranging between 12–38.3%. In addition, the maximum conversion obtained in this work is much higher than the previous results obtained by our research group with a fluidized bed reactor and a Cu/ γ -Al₂O₃ catalyst (a maximum conversion of 40% was reached) [40].

The CO₂ flowrate shows an effect on the energy efficiency of the process, as can be observed in Figure 4. The maximum CO₂ energy efficiency obtained in this study was 1.55%, using an Ar flowrate of 1 L/min and a CO₂ flowrate of 9 mL/min. This value was similar to the results obtained with a fluidized bed reactor (1.3-2.0%) [40]; however, still less efficient than other works published in the literature working with DBD reactors that obtained energy efficiencies ranging from 7 to 45.2% (Table 1). The main reason to explain these differences in energy efficiency is caused by the different operational conditions (use of diluent gas, CO₂ flowrate, catalyst, and power discharge). Probably, the amount of energy applied (power discharge) related to the CO₂ flowrate used is too high compared with the literature. When the CO₂ flowrate increased too much, the energy efficiency was affected by a lower CO₂ conversion, as observed previously in Figure 3.



Figure 3. Effect of CO_2 flowrate on the CO_2 conversion. Conditions: 1 L/min of Ar, 50 W, 20 kHz and minimum duty cycle.



Figure 4. Effect of CO_2 flowrate on the energy efficiency. Conditions: 1 L/min of Ar, 50 W, 20 kHz, and minimum duty cycle.

The formation of active species in the DBD reactor is entirely dependent on the initial concentration of reactants and the specific input energy (SIE). The SIE (kJ/L) was calculated with the following equation:

$$SIE = \frac{P_E \cdot 60}{F_{CO2}}$$
(4)

With a catalyst, not only the concentration of the reactants but also the reduced electric field enhances inside the plasma zone, which in turn generates more chemically reactive

species and significantly contributes to the activation of CO_2 . The conversion rate of CO_2 was determined from Equation (5) [16].

$$\ln\left(\frac{C_{\text{CO2, in}}}{C_{\text{CO2, out}}}\right) = \text{SIE} \cdot \mathbf{k} + \mathbf{c}$$
(5)

The conversion rate is expressed by k, and c stands for the intercept. Figure 5 reports the lineal increment of $\ln(C_{CO2}, in/C_{CO2}, out)$ as a function of SIE. The decomposition rate observed was of 0.0014 L/kJ. This value was lower than the value obtained by Ray et al. [43], who obtained a maximum conversion rate of 0.02889 L/kJ using 15% CuO/Al₂O₃ catalyst and without Ar dilution. They worked with lower SIE values which can also affect the conversion rate. So, it is plausible that the difference between conversion rates is caused by differences in the operational conditions.



Figure 5. Rate of reaction as a function of SIE (linear fit).

3.2. Influence of Argon Flowrate

In the previous section, it was revealed that at lower CO_2 flowrates the residence time in the DBD plasma reactor was higher, thus, improving CO_2 conversion. Another important variable of the process is the use of Ar as a diluent gas. For that reason, it was studied the influence of the Ar flowrate on the CO_2 conversion into CO. This study was also carried out at different CO_2 flowrates (so at different CO_2 concentrations), and the results are shown in Figure 6.

Figure 6 shows how the conversion increases linearly with increasing Ar flowrates as reported in other works [39,40]. During the experiments, it was observed that at higher Ar flowrates, the NTP generated was more uniform and with a higher number of discharges inside of the reactor. This explains why at a high Ar flowrate the conversions obtained are higher.



Figure 6. Effect of Ar flowrate on the CO_2 conversion. Conditions: 50 W, 20 kHz and minimum duty cycle. The experiments were performed at different CO_2 flowrates.

3.3. Environmental Impact Using Photovoltaic Energy to Power the DBD Reactor

The assessment of the ECco₂ was performed considering two different scenarios. The first scenario uses the electricity mix of Spain to run the DBD reactor, whereas the second scenario considers the current configuration of this work, supplying energy with a PV system. In Figure 7, it can be seen how the first scenario represents a lower compensation of CO₂ emissions (between 7–13%). It can be noticed that it shows a similar tendency to the energy efficiency plot (Figure 4), as they are correlated; therefore, with this first scenario, the emissions derived from the electricity mix are much higher than those which can be counterbalanced by the CO₂ captured and the CO produced during the DBD reaction. According to the Ecoinvent database, the emission factors related to the consumption of 1 kWh with the electricity mix of Spain and the production of 1 kg of industrial CO would be estimated as 0.131 kg of CO₂-eq and 1.51 kg of CO₂-eq, respectively.



Figure 7. Effect of CO₂ flowrate and the type of electricity source on the carbon emissions compensated. Conditions: 1 L/min of Ar, 50 W, 20 kHz, and minimum duty cycle.

The optimal experiment (highest energy efficiency) was at a CO₂ flowrate of 9 mL/min. In this experiment, the CO₂ converted to CO (C_{CO2}) was 0.43 g CO₂-eq/h, whereas the emission avoided by the production of CO (P_{CO}) was 0.41 g CO₂-eq/h. The energy consumption was 0.11 kWh per each gram of CO₂ transformed into CO. The emission related to the electricity consumption (E_{electricity}) considering the electricity mix of Spain was 6.42 g CO₂-eq/h.

Taking into account the PV system scenario, the ECco₂ rose up to 97% (Figure 7). This shows how the combination of the NTP process with PV systems significantly enhances the sustainability perspectives of the process. This increase would be justified by the lower emission factors related to the production of 1 kWh using PV panels (Ecoinvent database \rightarrow 0.018 kg of CO₂-eq). There is a difference of one order of magnitude lower than the emission factor of the electricity from the electrical grid. The emission related to the electricity consumption (E_{electricity}) in the optimal experiment (9 mL/min of CO₂) using the PV system was 0.86 g CO₂-eq/h. That is why the obtained results clearly evidenced that the combination of the DBD reactor with renewable energy significantly enhances the life cycle assessment of the process, reaching almost a complete compensation of the CO₂ emissions with a more sustainable perspective.

4. Conclusions

A cylindrical packed-bed DBD plasma reactor was used for the conversion of CO_2 into CO using a copper and zinc-based catalyst, copper electrodes, Ar as a diluent gas, and photovoltaic energy as an electricity source. Parameters such as CO_2 flowrate in the inlet gas stream and Ar flowrate were assessed. The main findings of this study are stated below:

- The concentration of CO₂ in the inlet of the reactor is an important variable. The CO₂ conversion is higher at lower CO₂ concentrations. An increase in CO₂ concentration causes a major decline in CO₂ conversion.
- The maximum CO₂ conversion was 74.2%, using an Ar flowrate of 1 L/min and a CO₂ flowrate of 3 mL/min, applying 50 W, a frequency of 20 kHz, and a minimum duty cycle.
- The presence of the diluent gas (Ar) has a strong influence on the decomposition of CO₂. It was observed that at higher Ar concentrations, the conversion improved.
- The use of photovoltaic energy increases the sustainability of the process. Using an LCA approach, it was estimated, for the decomposition reaction, that, with the best conditions obtained in this study, it would be possible to compensate 97% of the CO₂ emissions related to the process.

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